Naturalness is key: high species richness of woodinhabiting fungi does not automatically mean high species quality

Jan Holec $^{\rm l}$, Pavel Hubený $^{\rm 2}$ & Tomáš Kučera $^{\rm 3}$

1 Mycological Department, National Museum, Cirkusová 1740, Praha 9, CZ-193 00, Czech Republic

² Šumava National Park Administration, 1. máje 260, Vimperk, CZ-385 01, Czech Republic

 3 Department of Ecosystems Biology, Faculty of Science, University of South Bohemia, Branišovská 1760, České Budějovice, CZ-370 05, Czech Republic

e-mails: jan.holec@nm.cz; pavel.hubeny@npsumava.cz; kucert00@prf.jcu.cz

Holec J., Hubený P. & Kučera T. (2024): Naturalness is key: high species richness of wood-inhabiting fungi does not automatically mean high species quality. – Sydowia 76: 279–295.

Diversity of wood-inhabiting macrofungi on large decaying trunks of Norway spruce and Silver fir was monitored in Zámecký les near-natural forest in Czechia. The aim was to statistically evaluate the fungal species richness and composition in relation to environmental/trunk parameters and to compare it with data on forest naturalness taken from historical documents. The results were compared with data obtained by the same methods in Boubínský prales virgin forest and literature data from Mittelsteighütte natural forest. Surprisingly, trunks in the near-natural forest were species-richer than in the virgin one, showing that the available ecological niche, here a fallen trunk, can be occupied by a rich set of fungi regardless of human impact. However, species composition differed considerably among the sites, especially by the presence of rare, red-listed and old-growth forests fungi, designated as species of special interest (SSI). They were least represented in the near-natural forest, more in the natural forest, and most in the virgin forest. This correlation shows that the independent concepts of both SSI species and classification of forest naturalness go well together. Even seemingly small interventions in the past like selective cutting have a big impact on fungal communities. The most sensitive fungi like *Amylocystis lapponica*, *Fomitopsis rosea* or *Phellinus ferrugineofuscus* require unbroken forest continuity. They are absent from affected sites although their refugia as potential sources of propagules exist nearby. Our data document that only spruces and firs 500–600 years old indicate true forest continuity. Linking fungal occurrence data, environmental variables and historical documents on human interventions is crucial both for understanding ecosystem processes and conservation management.

Key words: Central Europe, Bohemian Forest, mixed montane forests, ecology, spread limitations.

Human pressure on the landscape, including forests, is constantly increasing. Mycologists are worried about its effect on diversity and composition of fungal communities in forests (Bengtsson et al. 2000, Lindner et al. 2006, Paillet et al. 2010, Blaser et al. 2013, Abrego & Salcedo 2013, Brazee et al. 2014, Heilmann-Clausen et al. 2014, Juutilainen et al. 2014, Purahong et al. 2014, Goldmann et al. 2015, de Groot et al. 2016, Suominen et al. 2019, Tomao et al. 2020, Heine et al. 2021). This applies in particular to wood-inhabiting fungi which lose their substrate by removing living and dead trees (Müller et al. 2007) and suffer from habitat changes connected with silvicultural practices (Bässler et al. 2010). Wood-inhabiting (lignicolous) fungi, functioning ecologically as endophytes, parasites, and decomposers, are ideal model groups, which enables answering general questions on species diversity, community ecology and conservation biology (Heilmann-Clausen & Christensen 2004; Küffer et al. 2008; Kubartová et al. 2012; Norros et al. 2012; Rajala et al. 2012, 2015; Abrego & Salcedo 2013; Ottosson et al. 2014, 2015; Hoppe et al. 2016; Juutilainen et al. 2017; Komonen & Müller 2018; Purhonen et al. 2019; Holec & Kučera 2020; Holec et al. 2020, 2022a; Moor et al. 2020; Nordén et al. 2020; Runnel et al. 2021; Abrego 2022; Rustøen et al. 2023).

Various studies have documented a reduction of the diversity of wood-inhabiting fungi in forests and a shift in their community composition caused mainly by these factors: clearcutting and selective logging (Bader et al. 1995, Lindblad 1998, Josefsson 2010), intensive thinning (Müller et al. 2007), salvage logging after insect outbreaks (Bässler et al. 2012), forest fragmentation (Abrego & Salcedo 2014, Grilli et al. 2017) and canopy gaps formation

(e.g. Perreault et al. 2023). Tomao et al. (2020) concluded that the higher is the forest management intensity the lower is the diversity of ectomycorrhizal and wood-inhabiting species, at least in the short term. Thus, the forest naturalness reflecting the degree of human interventions is a key factor determining especially the presence of endangered lignicolous species (Junninen et al. 2006). Some of them proved to be good indicators of forest naturalness (Bader et al. 1995, Kotiranta & Niemelä 1996, Parmasto 2001, Holec 2003, Christensen et al. 2004, Müller et al. 2007, Blaschke et al. 2009, Dvořák et al. 2017, Halme et al. 2017, Heilmann-Clausen et al. 2017). Presence of such fungi as well as high total mycodiversity is connected above all with the large forest size and tree cover continuity, plus rich presence and diversity of deadwood, especially of the huge living trees and big units of the coarse wood debris (CWD) (Paillet et al. 2010, Hofmeister et al. 2015, Ruokolainen et al. 2018, Runnel et al. 2021, Majdanová et al. 2023).

Although our knowledge of these phenomena is growing, it is surprising that studies on the impact of forest management on wood-inhabiting fungi rarely specify exact character and intensity of human impact. In most publications, only brief and vague information is given, such as "selectively logged in the past, partially deforested in the 19th

Fig. 1. Geographic location of the Zámecký les forest (ZL) and the localities used for comparison: Boubínský prales (BP) and Mittelsteighütte (MH). In figure b, notice the difference between the forest composition at the Zámecký les (large old trees, heterogeneous structure of the near-natural forest) and its surroundings towards the east and south (young trees, homogeneous structure of the man-made forest). Source of basic maps: Mapy.cz (www.mapy.cz), @Seznam.cz, a.s., 2023.

century" etc. As indirect evidence of forest continuity and past in terventions, the number of cut stumps per hectar was used by Bader et al. (1995) and tree age structure plus representation of oldest trees based on dendro chronology by Josefsson (2010) and Majdanová et al. (2023).

We decided to use historical documents to specify the time and type of human interventions in forest stands of different natu ralness. As a "virgin forest standard", we used Boubínský prales in Czechia, located in the Bohe mian Forest Mts. (Vrška et al. 2012, Holec et al. 2015). We com pared it with similarly looking stand in the same mountain range – Zámecký les forest near the village of Železná Ruda (Fig. 1). It is remarkable by multiaged structure with numerous huge individuals of spruce and fir, both living and dead, having a diameter of up to 130 cm, height up to 50 m and age up to 280 years. Just visually, the stand looks like a virgin forest (Fig. 2). However, it was significantly in fluenced by man in the past (Electronic Supplement A) as it was adjacent to the former De brník chateau (Schloss Deffernik, built in 1779, demolished in 1989) belonging to glassmaking fami lies (Fig. 3a). They used wood from surrounding forests for their glassworks already since 1774 and especially in the $19th$ century. Moreover, big and slowly decaying stumps after felling of large spruces and firs in the sec ond half of the 20th century are still visible, documenting selec tive cutting in not too distant his tory.

Fig. 2. Forest interior**. a.** Zámecký les forest (near-natural forest), **b.** Mittel steighütte (natural forest), **c.** Boubínský prales (virgin forest).

Fig. 3. Forest history at Zámecký les (ZL) documented by cadastre and forestry maps. **a.** State in 1837, original map of the stable cadastre. @ State Administration of Land Surveying and Cadastre of the Czech Republic, Archival maps, https://ags.cuzk.cz/archiv/openmap.html?typ=omc&idrastru=B2_a_4C_1122_6. Note the continuous forest (in grey) between the former Deffernik chateau and the former mirror factory (Spiegel Fabrik) in the Ferdinandsthal valley. **b.** State in 1876, archival forestry map, taken from Hubený (2023). The dark green colour of the ZL and some of the surrounding stands indicates forests older than 100 years, the grey colour of stands up to 20 years old. Note that the stands to the east and south of the ZL have been cleared shortly before 1876. c. State in 2016, current forestry map. Taken from Lesní hospodářská kniha (Forest management book), LHC 382216 – ÚP Prášily, 01.01.2016 – 31.12.2027 (depon. in Forests of the Czech Republic, state enterprise). The ZL stand having green colour is about 200 years old (average value, estimation) and red hatching indicates the existence of undergrowth with an age of around 40 years (composed mostly of beech).

We hypothesized that due to the aforementioned human interventions, the fallen decaying trunks in Zámecký les are species-poorer and less valuable in terms of the representation of rare, threatened and bioindicator species that in the Boubínský prales virgin forest. On the other hand, the Zámecký les mycobiota could have been re-enriched via propagules from a nearby refugium, the Mittelsteighütte nature reserve in the Bavarian Forest National Park, Germany. It is a very valuable natural forest of the similar tree composition and elevation like Zámecký les, located only 3.5 km away on the same-oriented slope in the Grosse Deffernik stream valley. Its rich and rare mycobiota is well known (Nuss 1999, DGfM 2023) and suitable for comparison.

Our questions were: 1. if the wood-inhabiting mycobiota of a stand that looks like a virgin forest, but has been principally affected by man (Zámecký les), is as species-rich and valuable like in the true virgin forest (Boubínský prales) and the nearby natural forest (Mittelsteighütte); 2. what are the main factors responsible for possible differences.

Materials and methods

Study site Zámecký les and the localities compared

All sites (see below) are located in the same mountain range, namely the Bohemian Forest on the border between the Czech Republic and Germany (Fig. 1). Data on their habitat conditions and forest history are summarized in Electronic Suppl. A. Vegetation of all sites is very similar and made up of herb-rich beech forest with a significant proportion of Norway spruce (*Picea abies*) and Silver fir (*Abies alba*) (Fig. 2). The classification of forest naturalness in Zámecký les, Boubínský prales and their surroundings was taken from the Czech Natural Forests Databank (https://naturalforests.cz/

czech-natural-forests-databank) under names Boubínský prales and NP Šumava - Medvědí jámy - Pod Sklářským vrchem. Data on Mittelsteighütte were taken from Nuss (1999) and classified according to the Proposal for terminology standardization at web page Naturalforests.cz (2023). In this document, three categories of the broadly conceived term natural forest (= old-growth forest) are recognized based on level of human interventions: 1. virgin (original) forest, 2. natural forest, 3. near-natural forest.

Zámecký les (ZL), near-natural forest (Figs. 2a, 3)

Czech Republic, located in the Bohemian Forest (= Šumava) National Park (NP), 1.6 km S of the village of Železná Ruda, site called Debrník (after former Debrník = Deffernik chateau), forest stand called "Zámecký les", natural zone of the NP, elevation 770–825 m a.s.l., coordinates of the centre 49.1230436N, 13.2352622E, area ± 8 ha, mixed montane forest (*Fagus sylvatica*: mostly younger and middle-aged trees; *Picea abies* and *Abies alba*: mostly large old trees) with multi-aged structure and high amount of coarse wood debris, habitat: herb-rich beech forest = *Asperulo-Fagetum* beech forest (Chytrý et al. 2010, AOPK ČR 2023). Data regarding human impact were obtained by P. Hubený from cadastre maps, historical and forestry archives (Fig. 3) and by J. Holec from the web site on glassworks in Debrník area (http://m.taggmanager.cz/ trail/cs/122). In short, the stand is a man-influenced successor of the original virgin forest currently surrounded by young and middle-aged managed spruce forests. Its complete deforestation was probably not permitted by the owners of the nearby Debrník chateau and glasswork (see Introduction).

Boubínský prales (BP), virgin forest (Fig. 2c)

Czech Republic, located in the Bohemian Forest (= Šumava) Protected Landscape Area. Habitat data were taken from Vrška et al. (2012) and Holec et al. (2015, 2020). The site is protected as a national nature reserve. It is a perfectly preserved original forest never affected by forestry interventions, very valuable even from a pan-European perspective.

Mittelsteighütte (MH), natural forest (Fig. 2b)

Germany, located in the Bavarian Forest National Park (= Nationalpark Bayerischer Wald), formerly protected as a nature reserve, now a strictly protected area ("Urwaldgebiet") inside the national park. Habitat data were taken from Nuss (1999). It

Studied trunks and their characteristics

The largest fallen trunks present in ZL were selected to be comparable with those from BP studied earlier and used here for comparison (Holec et al. 2020, Holec & Kučera 2020, Holec et al. 2022a). Their diameter at breast height was 90–130 cm for *Picea* and 85–115 cm for *Abies* (Electronic Suppl. C, D). As the number of such trunks was very limited, only 12 trunks of *Picea* (coded DP) and 6 of *Abies* (DA) could be selected (Electronic Suppl. B), covering all decay stages more or less equally. Their characteristics were recorded by J. Holec in August 2022 (Electronic Suppl. C, D): way of fall (broken or uprooted), diameter at breast height (DBH, in cm), length (m), geographic coordinates (using handheld Garmin GPSmap 60CSx device), direction of fall (in azimuth degrees), elevation (m a.s.l.), decay stage (1–5, average value for the entire trunk; estimated in accordance with Heilmann-Clausen 2001 and Holec et al. 2015, 2020), contact with the soil $(%)$, bark cover $(%)$, moss cover $(%)$, cover of trees (E3, %, estimated from a rectangle covering the trunk and 1 m more at both sides), cover of shrubs and young trees up to a height of 5 m (E2, %, estimated like E3), total canopy cover (E32, %). Trunk volume was calculated according to the formula for a truncated cone using ½ of the DBH as radius of the bottom disc and 2.5 cm as radius of the top disc (resulting from 5 cm as the usual width of the fallen trunk top). The trunk DA07 was represented only by its lower half having a diameter of the top disc 50 cm. In this case, 25 cm was used as radius of the top disc.

Monitoring of fungi in Zámecký les

Four mycological inspections were conducted on each trunk in ZL in 2021–2022, always at the time of the greatest fructification for the given period (spring visit: 15–16 June 2021, summer: 9–10 August 2021, late autumn: 15–16 November 2021, autumn: 29–30 September 2022). All visible macromycetes were recorded. For comparability, the field work and elaboration of fungal records was done in the same way as during our previous studies in Boubínský prales (Holec et al. 2020, 2022a; Holec & Kučera 2020). All fungi were recorded and identified by J. Holec. Most collections of polypores were revised by P. Vampola (Smrčná, Czechia), some collections of tomentelloid fungi by A. Jirsa (University of South Bohemia, Czechia). Vouchers of hardly identifiable taxa are deposited in the mycological herbarium of the National Museum, Prague (PRM 957149–957282, 959002–959099).

Data for comparison

Data from trunks monitored in BP by the same methods like in ZL were taken from our previous publications: 18 trunks of *Picea* studied in 2020 (Holec et al. 2022a) and 30 trunks of *Abies* studied in 2017–2019 (Holec & Kučera 2020), see Electronic Suppl. C, D (trunk characteristics) and Electronic Suppl. F, G (fungal occurrence data). In some cases, data from 33 trunks of *Picea* studied in 2015 (Holec et al. 2020) were also consulted. Data on fungi occurring in MH were taken from Nuss (1999) and DGfM database (DGfM 2023).

Species of special interest (SSI)

Threatened species included in the Czech Red List (Holec & Beran 2006), generally rare species and fungi preferring old-growth forests were classified as "species of special interest" (SSI). For the general concept of the SSI see e.g. Ódor et al. (2006). For selection of the old-growth forests species, see relevant references in introduction and the following publications focused on individual species: Antonín et al. (2011), Běťák et al. (2012, 2021), Holec & Kolařík (2017), Holec & Zehnálek (2021), Holec et al. (2019, 2022b, 2023), Vampola (2021), Langer et al. (2022).

Statistical evaluation

Explanatory trunk and habitat variables. Similarly to our previous studies (Holec et al. 2020, Holec & Kučera 2020), all explanatory variables were inspected for multicollinearity using principal component analysis (PCA, see Electronic Suppl. E). One group of the collinear trunk variables was represented by the bark cover (Bark), which was negatively correlated with decay stage (Decay, Pearson's correlation coefficient $r = -0.72$, $p < 0.001$), total canopy cover (E32, r= -0.53 , p < 0.001) and moss cover (Moss; $r = -0.50$, $p < 0.001$). Decay covered positively also the moss cover $(r = 0.61, p < 0.001)$, trunk contact with soil (Soil; $r = 0.57$, $p < 0.001$) and the total canopy cover $(r = 0.51, p < 0.001)$. Trunk dimension parameters were best represented by volume (Volume) being collinear with trunk length (Length; $r = 0.52$, $p < 0.001$) and the trunk diameter (DBH; $r = 0.78$, $p < 0.001$). On the other hand, the trunk length was negatively correlated with decay stage $(r = -0.49, p = 0.001)$. Of the trunk dimension parameters, only length had partial significant effect on species richness (Nspec, $r = 0.40$, $p < 0.001$). The total fungal species richness was tested to partial effect of locality and tree species using nested ANOVA. The canopy cover group of variables (E3, E2, E32) was best expressed by common effect of the trees and shrubs (E32), representing also the partial effects of both tree (E3; $r = 0.70$, $p < 0.001$) and shrub cover $(E2; r = 0.28, p = 0.03)$, respectively. Nevertheless, the tree and shrub covers were related negatively with each other $(r = -0.34, p = 0.005)$. The moss cover had significant relation to the total canopy cover $(r = 0.46, p < 0.001)$. The folded aspect of fallen trunk to south-west (FAsw), expressing its exposure to afternoon sun (and, thus, its drying out), was computed as described in Holec et al. (2019). As insignificant, it was not used in subsequent analyses.

Due to the high multicollinearity of the above mentioned explanatory variables (Electronic Suppl. E: right figure), we counted substitute variables derived from positions of trunks along the first and third PCA axes (PCAbark, PCA_DBH), whose advantage is orthogonality. PCAbark positively expressed the bark cover and negatively decay and contact with soil ($r = 0.85, -0.93$ and -0.62 , respectively; all with p < 0.001). PCA_DBH was positively related to the trunk length, DBH and volume $(r =$ 0.97, 0.97 and 0.87, respectively; all with p < 0.001). The derived variables were used in results (Fig. 6).

Species composition on trunks. The species occurrence matrix based on data from Electronic Suppl. F, G was studied using gradient analyses in the Canoco ver. 5.12 software (ter Braak & Šmilauer 2012) after deleting rare species occurring only on one trunk (singletons). The total length of the largest distance measured with detrended correspondence analysis (DCA) was 6.38, which allowed us to use the unimodal ordination methods. The relationship between species community pattern and the trunk/habitat variables was tested within a constrained ordination framework (canonical correspondence analysis, CCA) using a Monte Carlo permutation test (MCPT, number of permutations 4999). The explanatory effects of particular environmental variables were evaluated in MCPT with a stepwise procedure of selecting the significant variables, i.e. those having the best fit to species data. As ordinal values of decay stage were not strictly linear, we used the individual decay stages as supplementary nominal variables and projected them passively by the software into separate biplots (as they were insignificant) reflecting the same directions of ordinal space. The difference of the fungal community pattern between ZL and BP was tested using partial CCA with the factor of the host tree. For further explanations, see Šmilauer & Lepš (2014).

Results

Species richness

Zámecký les

We recorded 175 taxa of macrofungi on 18 trunks studied (Electronic Suppl. F, G). They represent 173 species plus 2 varieties of the same species (*Mycena epipterygia*). The high-frequent species are listed in Electronic Suppl. H, I. None of the species was found on all trunks. Most species were basidiomycetes (168 of 175). The species-richest genera were *Mycena* (13 species), *Galerina* (7), *Botryobasidium* (6), *Xylodon* (5), *Hyphoderma* (4), and *Oligoporus* (4). Macroscopic ascomycetes were represented by only 7 species belonging to discomycetes (*Ascocoryne cylichnium*, *A. sarcoides*) and pyrenomycetes (*Camarops tubulina*, *Durandiella gallica*, *Echinosphaeria canescens*, *Hypocrea citrina*, *H. rufa*). Only *C. tubulina* and *A. cylichnium* were more frequent (5 trunks each).

Picea trunks (12) were inhabited by 142 species with 21–39 species per trunk (Electronic Suppl. F, Tab. 1). The average number of species per trunk was 30. The most frequent species were *Fomitopsis pinicola*, *Mycena rubromarginata*, *Xylodon asper* and *Physisporinus sanguinolentus*. Almost two thirds of the species $(92, i.e. 65%)$ were low frequent, found only on one trunk (singletons: 65) and two trunks (doubletons: 27).

Abies trunks (6) were inhabited by 104 species with 23–40 species per trunk (Electronic Suppl. G, Tab. 1). The average number of species per trunk was 32. The most frequent species were *Athelia epiphylla*, *Pluteus pouzarianus* and *Xylodon brevisetus*. Almost three quarters of the species (76, i.e. 73 %) were low frequent, found only on one trunk (singletons: 59) and two trunks (doubletons: 17).

Zámecký les versus Boubínský prales

A total of 307 species were found at both locations (Electronic Suppl. F, G; Tab. 2) on the 66 trunks studied. As shown in Tab. 1 and Fig. 4, no significant difference in fungal richness between individual *Picea* and *Abies* trunks at the same locality was detected, while the effect of locality on a specific tree species was significant (ANOVA F value = 13.4, $p =$ 0.0005). On average, the trunks of both *Picea* and *Abies* were species-richer in ZL than in BP, especially in the case of *Abies*. One outlayer (Fig. 4: at the top right) was the extremely rich *Picea* trunk BB13 (55 species).

Tab. 1. Number of species on *Picea* and *Abies* trunks at Zámecký les and Boubínský prales localities. For simplicity, the two varieties of *Mycena epipterygia* (Electronic Suppl. F, G) are counted as two species.

Tab. 2. Species occurring on *Picea* and *Abies* trunks at Zámecký les and Boubínský prales localities. For simplicity, the two varieties of *Mycena epipterygia* (Electronic Suppl. F, G) are counted as two species.

Fig. 4. Range of fungal species richness on all studied trunks in relation to the tree species and locality (BP: Boubínský prales, ZL: Zámecký les). Only the difference between localities was significant (ANOVA F value = 13.4 , $p = 0.0005$).

Species composition

Zámecký les

Fungi occurring on both *Picea* and *Abies* trunks represented only 40.5 % of the total species number (Tab. 2). The remaining species were unique to either *Picea* or *Abies* (40.5 %, 19 %; respectively). This shows that the fungal species composition on *Picea* and *Abies* trunks was quite different.

Zámecký les versus Boubínský prales

Combined data for ZL and BP showed a similar percentage of shared *Picea* + *Abies* species (42 %) as in the ZL itself (see previous section). However, the percentage of species unique to either *Picea* or *Abies* (26 %, 32 %; respectively) was reversed in favor of *Abies* (Tab. 2). This is mainly associated with the big increase of *Abies* species from 30 trunks studied at BP (compare only 6 available trunks at ZL).

If analyzed with multivariate statistical methods (Fig. 5), the species composition of fungal communities on individual trunks was clearly different between ZL and BP, both for *Picea* and *Abies*. Trunks from these two sites formed two separate clusters without any mixing (Fig. 5: right diagrams). Left diagrams of Fig. 5 document again that trunks in ZL are species richer (compare previous section and Fig. 4), a fact even more evident when singletones are eliminated.

The species composition of fungal communities on both *Picea* and *Abies* trunks (Fig. 6: top diagrams) was closely associated with the way of tree fall (uprooted or broken) and two derived variables: PCAbark and PCA_DBH. PCAbark expresses collinear variables bark cover, decay, moss cover, contact with soil and the total canopy cover (see Electronic Suppl. E; note that some of them are correlated negatively, i.e. in opposite direction). PCA_ DBH represents the trunk volume parameters (trunk length, DBH, volume). When wood decay stages are passively projected into space defined by these variables, a slightly different trajectory of wood decay for different tree is seen (Fig 6: bottom diagrams) – a curve resembling horizontal letter S for *Picea* and letter U for *Abies*. It shows that the group of species corresponding to the later decay stages 4 and 5 is more distinct in *Abies*, consisting of some mycorrhizal species (*Laccaria amethystina, Lactarius subdulcis*) and corticioids (e.g. *Leptosporomyces galzinii, Athelia decipiens, Hyphoderma cremeoalbum, Botryobasidium subcoronatum*).

Tab. 3. Presence of species of special interest (SSI) in Zámecký les and Boubínský prales.

Fig. 5. Ordination biplots (CCA) defined by fungal species patterns for *Picea* and *Abies*. **Left** are isolines showing gradient of fungal richness on trunks at projected localities (BP: Boubínský prales, ZL: Zámecký les). **Right** are centroid positions of fungal communities on the individual studied trunks grouped by locality. For *Picea* and *Abies*, first and second axes explain 18.3 % and 18.8 % of fungal species variation, respectively. For trunk codes (BB, DP, BA, DA), see Electronic Suppl. C, D. Species numbers and community compositions were computed without singletones.

Species of special interest (SSI)

SSI species are less represented in ZL than in BP, both on *Picea* and especially on *Abies* trunks (data: Electronic Suppl. F, G; summarization: Tab. 3). The percentage of SSI species common to both localities (Tab. 3) is relatively low (25 % for *Picea*, 19 % for *Abies*) which means that most SSI are different from one locality to another. When only the most valuable species are highlighted, i.e. the rarest ones and those preferring old-growth forests, the difference between ZL and BP is even greater (Tab. 4). Boubínský prales is much richer in these species, hosting rarities like *Amylocystis lapponica*, *Chrysomphalina grossula*, *Clitocybula familia*, *Cystoderma subvinaceum*, *Dentipratulum bialoviesense*, *Fomitopsis rosea*, *Galerina pruinatipes*, *Ionomidotis irregularis*, *Junghuhnia collabens*, *Kneiffiella altaica*, *K. curvispora*, *Laurilia sulcata*, *Mycena clavata*, *Phellinus pouzarii, Pseudoplectania melaena, Resupinatus striatulus* and *Skeletocutis odora* (Tab. 4).

Discussion

Environmental variables shaping fungal richness and species composition on decaying trunks of *Picea* and *Abies* have been thoroughly discussed in our previous papers from BP locality (Holec et al. 2020 , $2022a$; Holec & Kučera 2020). In agreement with publications cited there, the key role of wood decay stage with associated variables bark and moss cover was shown. Other significant variables are the canopy cover (influencing heat load, amount of precipitation and air humidity), elevation, tree history (expressed by time since death, time since fall and way of fall) and trunk volume. New data from ZL further confirm general importance of these parameters (Electronic Suppl. E, Fig. 6). In the following sections, we will focus on the influence of forest naturalness.

Regarding the relevance of our fruitbodiesbased research in the molecular era, we discussed it in detail in previous papers (Holec et al. 2020, Holec & Kučera 2020) showing that this method is not outdated. We agree with the latest synthesis published by Heine et al. (2021). They concluded that eDNA metabarcoding cannot be used interchangeably for morphological community analyses to identify response patterns of fungal communities on forest management strategies. The best way is a combination of both approaches. This is a challenge for future research at our study sites.

Species richness

The fact that the trunks of both *Picea* and especially *Abies* are on average species-richer in forest affected by man (ZL) than in the virgin one (BP) is quite surprising (Fig. 4). We would assume the opposite, especially with regard to the huge species pool and dead wood supply in BP (Holec et al. 2015, Vrška et al. 2012), a site of much larger area (Electronic Suppl. A). It turns out, however, that a certain ecological niche, here a fallen decaying trunk, can be inhabited by a rich community of fungi regardless of the degree of human intervention on the site. It is also likely that the lower elevation of ZL (one of the key factors, see previous section) enables occurrence of higher number of species in comparison with higher-located (= cooler) BP site, especially for *Abies*. The unusual fungal richness of one *Picea* trunk from BP (Fig. 4, Electronic Suppl. F: BB13) is caused by the presence of an extraordinary high number of old-growth forests species that are absent from ZL (Tab. 4).

Most publications simply state that fungal diversity in man-influenced forests is lower than in the natural stands (e.g. Bader et al. 1995, Lindblad 1998, Junninen et al. 2006, Müller et al. 2007, Tomao et al. 2020). However, it concerns the total diversity at study site or plot. Our results show that the discrete substrate units in man-influenced forest may be inhabited by equally rich or even richer fungal communities than in the natural ones. A similar situation was documented by Suominen et al. (2019) based on fruitbodies-based research and Purahong et al. (2014) by molecular methods concluding that "forest management may affect fungal OTU richness in CWD logs in a complex manner".

However, there is a substantial difference in species composition between trunks in ZL and BP (Fig. 5), which is discussed below.

Species composition

Both in ZL and BP, *Picea* and *Abies* occur together. It would be expected that the dead trunks of these two conifers would host similar fungal communities. However, the difference is quite big. In ZL, the species unique to either *Picea* or *Abies* represent almost 60 % of the total species number and the percentage is almost the same when ZL and BP are counted together (Tab. 2: 58 %). This shows that many species of lignicolous fungi have a relatively strong preference for the tree species they inhabit. This fact was recently stressed e.g. by Rustøen et al. (2023), although in their case the difference especially between deciduous and coniferous trees was

Tab. 4. Comparison of presence of species of special interest (SSI) between Zámecký les (ZL), Mittelsteighütte (MH) and Boubínský prales (BP). Only the rarest and old-growth forest species are shown. See Materials and methods for sources of data, delimitation of SSI species and categories of naturalness, and Electronic Suppl. F, G for presentation of all SSI species.

A on *Abies* trunk

P on *Picea* trunk

- § known from Mittelsteighütte according to Nuss (1999) or DGfM (2023)
- & recently described, less known species
- species known outside monitored trunks (see Holec et al. 2015)

** according to Nuss (1999)

*** according to DGfM (2023)

shown. Our results document that the difference is distinct even between conifers and at the same site. It is certainly also connected with the different properties of spruce and fir wood, where fir wood is softer, without resin, and therefore rots more easily.

The difference between localities is also high. Fungal communities on trunks in ZL and BP are clearly separated for both *Picea* and *Abies* (Fig. 5). This documents a large influence of factors associated with the locality. As shown in Electronic Suppl. A, the main differing parameters between ZL and BP are the altitude (already discussed above), area and degree of forest naturalness. The last two factors are discussed in the following sections.

Species of special interest (SSI)

This group of species contains not only rare and/ or endangered Red List fungi, but also species demonstrably preferring old-growth forests ("indicator species": Bader et al. 1995, Kotiranta & Niemelä

1996, Parmasto 2001, Holec 2003, Christensen et al. 2004, Müller et al. 2007, Blaschke et al. 2009, Dvořák et al. 2017, Halme et al. 2017). SSI species are much less represented in ZL than in BP (Tabs. 3, 4). The most valuable old-growth forest species are totally absent in ZL (see results and Tab. 4; especially *Amylocystis lapponica, Phellinus pouzarii, Pseudoplectania melaena, Clitocybula familia*), not only from the monitored trunks but also from the whole locality (own data, unpublished). On the other hand, several other valuable species are known from BP outside the studied trunks, namely *Antrodia cretacea*, *A. piceata*, *Chrysomphalina chrysophylla* and *Phellinus ferrugineofuscus* (Tab. 4). Most of the exclusive ZL fungi (Tab. 4) are represented by recently described inconspicuous species that are likely to be found in BP during a future survey. The Mittelsteighütte reserve is somewhere between ZL and BP (Tab. 4). Generally, the fungal occurrence data from all three sites show a perfect match between representation of SSI species and the degree of forest naturalness: virgin forest for BP, natural forest for MH, and near-natural forest for ZL. It documents that the independent concepts of SSI species (as defined in Materials and methods) and classification of forest naturalness (according to Naturalforests 2023) go well together.

Although ZL belongs to the "worst category" of natural forests, it still hosts some old-growth forests fungi (Electronic Suppl. F, G). They are less rare than species listed in Tab. 4, but still worth mentioning: *Alutaceodontia alutacea, Antrodiella citrinella, Baeospora myriadophylla, Botryobasidium medium, Callistosporium pinicola, Camarops tubulina, Clitocybula lacerata, Crepidotus kubickae, Galerina stordalii, Gymnopilus bellulus, Hericium flagellum, Hyphoderma capitatum, H. obtusiforme, Kneiffiella floccosa, K. cineracea, Lentinellus castoreus, Mycena laevigata, Panellus violaceofulvus, Phellinus nigrolimitatus, Rigidoporus crocatus, Steccherinum gracile*. The locality is therefore definitely valuable as regards the wood-inhabiting fungi. The question remains why ZL is not even \P , better" when it has a nearby source of mycelia and spores of many other old-growth forest fungi, namely the Mittelsteighütte reserve (Fig. 1c, Tab. 4). It is discussed in the next section.

Naturalness and size, key factors

Based on our data, we agree with Bässler et al. (2010) that silvicultural strategies are very important for preservation of wood-inhabiting fungi in forests, as they influence two key factors responsible for their diversity: 1) forest naturalness and 2) amount and variety of available resource, i.e. dead wood (Abrego & Salcedo 2013). In general, rich presence of CWD was shown to be essential for high species richness and presence of red-listed species (Juutilainen et al. 2014). The dead wood of large volume proved to be most important (Runnel et al. 2021). Hofmeister et al. (2015) stressed the importance of trunks having diameter >80 cm. All trunks at ZL and BP studied by us were from this size class, however, they differed in fungal richness and composition between the sites. The decisive factor proved to be the degree of naturalness reflecting intensity of past human interventions, especially for the representation of SSI species. The main study site ZL is not just a mere near-natural forest but also a small "island" (about 8 ha) surrounded by the "sea" of more or less managed spruce forests. The first phase of selective logging took place already in the 18th century (Electronic Suppl. A). The biggest old trees were cut down, because the current oldest trees are only around 280 years old (compare BP with uninterrupted history, where trees are up to 600 years old, Electronic Suppl. A). The supply of dead wood of old trees was therefore interrupted at ZL, which today is shown in the absence of the most sensitive old-growth forest fungi (Tab. 4).

The importance of the decaying logs continuity was stressed e.g. by Bader et al. (1995) showing that selective logging about 100 years ago significantly decreased both the total species number and the number of threatened species. As documented by Josefsson et al. (2010), just minor forest logging (22– 26 cut stumps per hectar) carried out a century ago may have continuing effects on forest characteristics, including dead wood dynamics and the woodinhabiting fungal community, especially the abundance of red-listed species. Recently, Majdanová et al. (2023) confirmed that forest continuity indicated by the presence of >250 years-old-trees (i.e. those that survived logging and wood extraction around the end of the 18th century, similarly like in ZL site) is crucial for high richness of red-listed species. Our data from BP (Tab. 4) show that only spruces and firs 500–600 years old indicate a true continuity undisturbed by human interventions. Only such trees reach the maximal possible age (for spruce and fir) and then die naturally, providing fungi with the largest possible volume of dead wood. The data from our study and the mentioned papers complement each other nicely. It seems that the most sensitive fungi (those printed in bold in Tab. 4) require completely unbroken continuity and absent from even minimally affected sites.

Fig. 6. Ordination biplots (CCA) of fungal species composition on trunks of *Picea* and *Abies* from Zámecký les and Boubínský prales (**top diagrams**). The displayed significant variables (PCAbark; PCA_DBH; way of fall with two values: broken, uprooted) explain 17.4 % of total species variance for *Picea* (F-ratio = 1.8, p value < 0.001) and 13.9 % for *Abies* (F-ratio = 1.7, p value < 0.001). The first two axes cover 14.4 % and 11.7 % of cumulative fungal variance, respectively. For full names of the fungi, see Electronic Suppl. F, G. Singletones were excluded from analyses. Only fifty fungal species with higher fit were visualized. In **bottom diagrams**, the categories of decay (D1–D5) are projected passively (see Materials and methods). These diagrams relate to the upper ones and mirror the same quadrants of the biplots.

The small area is another unfavourable factor of the ZL site. The negative effect of forest fragmentation and interrupted connectivity is well known (e.g. Abrego & Salcedo 2014, Abrego et al. 2015, Grilli et al. 2017). Our comparison of ZL and MH, locations situated 3.5 km away, is in line with Abrego et al. (2015) who stressed the importance of connectivity at both site level (= locality area) and the regional scale of 10 km. Even if MH seems to be a perfect pool for recolonization of ZL by rare old-growth forest fungi, the reality is different as the ZL site lacks many valuable old-growth forest species (Tab. 4). Possible causes are discussed in the following section.

Limits to the spread of fungi in old-growth forests

Generally, the pool of available species is very important, being naturally greater in well-preserved and large stands. In them, fungi occurring on

old decaying trunks can gradually colonize newly fallen trunks in their vicinity (Lindblad 1998, Jönsson et al. 2008), that is, over a short distance. Recent studies revealed the dispersal limitation of specialist species already at distances of tens to a few hundred meters from the nearest fruitbody (Norros et al. 2012). It was also shown that most basidiomycete spores fall within 1 m of the cap (Galante et al. 2011), limiting their large-scale dispersal. Our data (especially the comparison of ZL and MH sites in Tab. 4) show that the rare old-growth forest fungi are limited in their ability to spread over several kilometers. On the other hand, the role of wood-inhabiting beetles carrying a broad range of woodinhabiting fungi and contributing to their dispersal was discovered recently (Seibold et al. 2019, Lunde et al. 2023). All these factors certainly affect the presence/absence of fungi at our sites.

As summarized by Abrego (2022), the ecological assembly processes are regulated by combination of environmental filters (factors preventing or facilitating species colonization and persistence), biotic filters (intra- and interspecific interactions), dispersal (movement and migration) and stochastic processes (random events, especially the priority effects; see e.g. Ottosson et al. 2014, Peay & Bruns 2014). In addition, the role of colonization–extinction dynamics was recently stressed by Moor et al. (2020) who showed that increasing specialization of some species is associated with increasing sensitivity to habitat conditions and increasing extinction rates. On the other hand, the colonization probability increases with larger numbers of suitably large logs in the right decay stage. Thus, forest age and total dead-wood volume are the main variables explaining the colonization probability of wood-decaying fungi (Moor et al. 2020). Interestingly, all species evaluated by Moor et al. (2020) are present in BP, but those of them where the highest degree of specialization has been shown (*Amylocystis lapponica, Fomitopsis rosea*, *Phellinus ferrugineofuscus*), are absent in ZL. This is surely connected with the scarcity of suitable substrates in ZL, a small site with partly interrupted continuity (Electronic Suppl. A).

Concluding remarks

Data on the diversity of wood-inhabiting fungi from sites of different naturalness show the importance of linking fungal occurrence data, environmental variables (obtained by field research) and historical data on human interventions in forests (obtained from archival documents, now increasingly available online). The last-named source enables reliable documentation of effects that biologists and ecologists can only capture indirectly or not at all. In our case, the historical data clearly showed the past human interventions, which today are negatively reflected in the representation of rare and old-growth forests fungi. It is a big task for nature conservation in protected areas to minimize the human influence and let the natural processes run, which could support the future return of sensitive species from surrounding refugia, if they still exist.

Acknowledgements

We thank P. Vampola (Smrčná, Czech Republic) for revision of some polypore collections and A. Jirsa (University of South Bohemia, České Budějovice, Czech Republic) for revision of some *Tomentella* collections. The work of J. Holec was financially supported by the Ministry of Culture of the Czech Republic (DKRVO 2024–2028/3.I.a, 00023272).

References

- Abrego N. (2022) Wood-inhabiting fungal communities: Opportunities for integration of empirical and theoretical community ecology. *Fungal Ecology* **59**, 101112: 1–8.
- Abrego N., Salcedo I. (2013) Variety of woody debris as the factor influencing wood-inhabiting fungal richness and assemblages: Is it a question of quantity or quality? *Forest Ecology and Management* **291**: 377–385.
- Abrego N., Salcedo I. (2014) Response of wood-inhabiting fungal community to fragmentation in a beech forest landscape. *Fungal Ecology* **8**: 18–27.
- Abrego N., Bässler C., Christensen M., Heilmann-Clausen J. (2015) Implications of reserve size and forest connectivity for the conservation of wood-inhabiting fungi in Europe. *Biological Conservation* **191**: 469–477.
- Antonín V., Beran M., Borovička J., Dvořák D., Holec J. (2011) *Clitocybula familia* (Fungi, Agaricales) – taxonomy, distribution, ecology and first records in the Czech Republic and Slovakia. *Czech Mycology* **63**: 1–11.
- AOPK ČR (2023) *Mapování biotopů (Mapping of habitats)*; https://aopkcr.maps.arcgis.com/apps/webappviewer/in dex.html?id=c38db59779714a78aec4c731152b0290 (accessed 9 Feb 2023).
- Bader P., Jansson S., Jonsson B.G. (1995) Wood-inhabiting fungi and substratum decline in selectively logged boreal spruce forests. *Biological Conservation* **72**: 355–362.
- Bässler C., Müller J., Dziock F., Brandl R. (2010) Effects of resource availability and climate on the diversity of wooddecaying fungi. *Journal of Ecology* **98**: 822–832.
- Bässler C., Müller J., Svoboda M., Lepšová A., Hahn C., Holzer H., Pouska V. (2012) Diversity of wood-decaying fungi under different disturbance regimes - a case study from spruce mountain forests. *Biodiversity and Conservation* **21**: 33–49.
- Bengtsson J., Nilsson S.G., Franc A., Menozzi P. (2000) Biodiversity, disturbances, ecosystem function and management of European forests. *Forest Ecology and Management* **132**: 39–50.
- Běťák J., Holec J., Beran M., Riebesehl J. (2021) Ecology and distribution of *Kneiffiella curvispora* (Hymenochaetales, Basidiomycota) in Central Europe and its phylogenetic placement. *Nova Hedwigia* **113**: 161–189.
- Běťák J., Pärtel K., Kříž M. (2012) *Ionomidotis irregularis* (Ascomycota, Helotiales) in the Czech Republic with comments on its distribution and ecology in Europe. *Czech Mycology* **64**: 79–92.
- Blaschke M., Helfer W., Ostrow H., Hahn C., Loy H., Bußler H., Krieglsteiner L. (2009) Naturnähezeiger – Holzbewohnende Pilze als Indikatoren für Strukturqualität im Wald. *Natur und Landschaft* **84**: 560–566.
- Blaser S., Prati D., Senn-Irlet B., Fischer M. (2013) Effects of forest management on the diversity of deadwood-inhabiting fungi in Central European forests. *Forest Ecology and Management* **304**: 42–48.
- Brazee N.J., Lindner D.L., D'Amato A.W., Fraver S., Forrester J.A., Mladenoff D.J. (2014) Disturbance and diversity of wood-inhabiting fungi: effects of canopy gaps and downed woody debris. *Biodiversity and Conservation* **23**: 2155– 2172.
- Christensen M., Heilmann-Clausen J., Walleyn R., Adamčík S. (2004) *Wood-inhabiting fungi as indicators of nature value in European beech forests*. In: *Monitoring and indicators of forest biodiversity in Europe – from ideas to operationality* (ed. Marchetti M.) European Forest Institute, Joensuu, EFI Proceedings No. **51**: 229–237.
- Chytrý M., Kučera T., Kočí M., Grulich V., Lustyk P. (2010) *Katalog biotopu*˚ *Cˇ eské republiky (Habitat catalogue of the Czech Republic)*. AOPK CR, Praha.
- de Groot M., Eler K., Flajšman K., Grebenc T., Marinšek A., Kutnar L. (2016) Differential short-term response of functional groups to a change in forest management in a temperate forest. *Forest Ecology and Management* **376**: 256– 264.
- DGfM (2023) *Datenbank der Pilze Deutschlands, Deutsche Gesellschaft für Mykologie e. V., Bearbeitet von Dämmrich F., Gminder A., Hardtke H.-J., Karasch P., Schmidt M.*; http://www.pilze-deutschland.de (accessed 6 Mar 2023).
- Dvořák D., Vašutová M., Hofmeister J., Beran M., Hošek J., Běťák J., Burel J., Deckerová H. (2017) Macrofungal diversity patterns in central European forests affirm the key importance of old-growth forests. *Fungal Ecology* **27**: 145–154.
- Galante T.E., Horton T.R., Swaney D.P. (2011) 95 % of basidiospores fall within 1 m of the cap: a field-and modelingbased study. *Mycologia* **103**: 1175–1183.
- Goldmann K., Schöning I., Buscot F., Wubet T. (2015) Forest management type influences diversity and community composition of soil fungi across temperate forest ecosystems. *Frontiers in Microbiology* **6**, 1300: 1–11.
- Grilli G., Longo S., Huais P.Y., Pereyra M., Verga E.G., Urcelay C., Galetto L. (2017) Fungal diversity at fragmented landscapes: synthesis and future perspectives. *Current Opinion in Microbiology* **37**: 161–165.
- Halme P., Holec J., Heilmann-Clausen J. (2017) The history and future of fungi as biodiversity surrogates in forests. *Fungal Ecology* **27**: 193–201.
- Heilmann-Clausen J. (2001) A gradient analysis of communities of macrofungi and slime moulds on decaying beech logs. *Mycological Research* **105**: 575–596.
- Heilmann-Clausen J., Adamčík S., Bässler C., Halme P., Krisai-Greilhuber I., Holec J. (2017) State of the art and future directions for mycological research in old-growth forests. *Fungal Ecology* **27**: 141–144.
- Heilmann-Clausen J., Aude E., van Dort K., Christensen M., Piltaver A., Veerkamp M., Walleyn R., Siller I., Standovár T., Ódor P. (2014) Communities of wood-inhabiting bryophytes and fungi on dead beech logs in Europe - reflecting substrate quality or shaped by climate and forest conditions? *Journal of Biogeography* **41**: 2269–2282.
- Heilmann-Clausen J., Christensen M. (2004) Does size matter? On the importance of various dead wood fractions for fungal diversity in Danish beech forests. *Forest Ecology and Management* **201**: 105–117.
- Heine P., Hausen J., Ottermanns R., Roß-Nickoll M. (2021) Comparing eDNA metabarcoding with morphological analyses: Fungal species richness and community composition of differently managed stages along a forest conversion of Norway spruce towards European beech in Germany. *Forest Ecology and Management* **496**, 119429: 1–13.
- Hofmeister J., Hošek J., Brabec M., Dvořák D., Beran M., Deckerová H., Burel J., Kříž M., Borovička J., Běťák J., Vašutová M., Malíček J., Palice Z., Syrovátková L., Steinová J., Černajová I., Holá E., Novozámská E., Čížek L., Iarema V., Baltaziuk K., Svoboda T. (2015) Value of old forest attributes related to cryptogam species richness in temperate

forests: A quantitative assessment. *Ecological Indicators* **57**: 497–504.

- Holec J. (2003) Auf natürliche, vom Menschen nur minimal beeinflusste Vegetation beschränkte Großpilze. *Fritschiana* **42**: 25–27.
- Holec J., Beran M., eds (2006) *Cˇervený seznam hub (makromycetu*˚*) Cˇeské republiky (Red list of fungi (macromycetes) of the Czech Republic)*. Příroda, Praha, 24: 1-282.
- Holec J., Běťák J., Dvořák D., Kříž M., Kuchaříková M., Krzyściak-Kosińska R., Kučera T. (2019) Macrofungi on fallen oak trunks in the Białowieża Virgin Forest - ecological role of trunk parameters and surrounding vegetation. *Czech Mycology* **71**: 65–89.
- Holec J., Dvořák D., Zíbarová L., Beran M., Zehnálek P., Peiger M., Kunca V. (2023) *Mycena laevigata* (Fungi, Agaricales) in the heart of Central Europe – a prominent species of old-growth forests. *Czech Mycology* **75**: 35–52.
- Holec J., Holcová K., Žák M. (2022a) Diversity and ecology of macrofungi on large decaying spruce trunks: what has changed after five years? *Sydowia* **75**: 23–35.
- Holec J., Kolařík M. (2017) First report of *Mycena clavata* (Fungi, Agaricales) in the Czech Republic including notes on its taxonomy, phylogenetic position and ecology. *Czech Mycology* **69**: 1–14.
- Holec J., Kříž M., Pouzar Z., Šandová M. (2015) Boubínský prales virgin forest, a Central European refugium of boreal-montane and old-growth forest fungi. *Czech Mycology* **67**: 157–226.
- Holec J., Kučera T. (2020) Richness and composition of macrofungi on large decaying trees in a Central European oldgrowth forest: a case study on silver fir (*Abies alba*). *Mycological Progress* **19**: 1429–1443.
- Holec J., Kučera T., Běťák J., Hort L. (2020) Macrofungi on large decaying spruce trunks in a Central European oldgrowth forest: what factors affect their species richness and composition? *Mycological Progress* **19**: 53–66.
- Holec J., Kunca V., Krˇíž M., Zehnálek P. (2022b) *Cyphella digitalis* (Fungi, Agaricales) – new data on ITS barcode, ecology and distribution in the Czech Republic and Slovakia. *Czech Mycology* **74**: 77–92.
- Holec J., Kunca V., Vampola P., Beran M. (2019) Where to look for basidiomata of *Phellinidium pouzarii*, a rare European polypore of montane old-growth forests with fir (*Abies*)? *Nova Hedwigia* **109**: 379–397.
- Holec J., Zehnálek P. (2021) Remarks on taxonomy and ecology of *Dentipratulum bialoviesense* based on records from Boubínský prales virgin forest in the Czech Republic. *Czech Mycology* **73**: 121–135.
- Hoppe B., Purahong W., Wubet T., Kahl T., Bauhus J., Arnstadt T., Hofrichter M., Buscot F., Krüger D. (2016) Linking molecular deadwood-inhabiting fungal diversity and community dynamics to ecosystem functions and processes in Central European forests. *Fungal Diversity* **77**: 367–379.
- Hubený P. (2023) *Atlas historických lesních map Šumavy (Atlas of historical forest maps of Šumava)*. Správa Národního parku Šumava (Administration of the Šumava National Park), Vimperk.
- Josefsson T., Olsson J., Östlund L. (2010) Linking forest history and conservation efforts: effects of logging on forest structure and diversity of wood-inhabiting fungi. *Biological Conservation* **143**: 1803–1811.
- Jönsson M.T., Edman M., Jonsson B.G. (2008) Colonization and extinction patterns of wood-decaying fungi in a boreal old-growth *Picea abies* forest. *Journal of Ecology* **96**: 1065–1075.
- Junninen K., Similä M., Kouki J., Kotiranta H. (2006) Assemblages of wood-inhabiting fungi along the gradients of succession and naturalness in boreal pine-dominated forests in Fennoscandia. *Ecography* **29**: 75–83.
- Juutilainen K., Mönkkönen M., Kotiranta H., Halme P. (2014) The effects of forest management on wood-inhabiting fungi occupying dead wood of different diameter fractions. *Forest Ecology and Management* **313**: 283–291.
- Juutilainen K., Mönkkönen M., Kotiranta H., Halme P. (2017) Resource use of wood-inhabiting fungi in different boreal forest types. *Fungal Ecology* **27**: 96–106.
- Komonen A., Müller J. (2018. Dispersal ecology of deadwood organisms and connectivity conservation. *Conservation Biology* **32**: 535–545.
- Kotiranta H., Niemelä T. (1996) *Uhanalaiset käävät Suomessa (Threatened polypores of Finland)*. 2nd revised edn. Edita, Helsinki.
- Kubartová A., Ottosson E., Dahlberg A., Stenlid J. (2012) Patterns of fungal communities among and within decaying logs, revealed by 454 sequencing. *Molecular Ecology* **21**: 4514–4532.
- Küffer N., Gillet F., Senn-Irlet B., Aragno M., Job D. (2008) Ecological determinants of fungal diversity on dead wood in European forests. *Fungal Diversity* **30**: 83–95.
- Langer E., Běťák J., Holec J., Klug A., Riebesehl J. (2022) *Kneiffiella altaica* and *Kneiffiella subaltaica* sp. nov. – a rare species found in European old-growth forests and a closely related new species from North America. *Nova Hedwigia* **115**: 205–225.
- Lindblad I. (1998) Wood-inhabiting fungi on fallen logs of Norway spruce: relations to forest management and substrate quality. *Nordic Journal of Botany* **18**: 243–255.
- Lindner D.L., Burdsall H.H., Stanosz G.R. (2006) Species diversity of polyporoid and corticioid fungi in northern hardwood forests with differing management histories. *Mycologia* **98**: 195–217.
- Lunde L.F., Boddy L., Sverdrup-Thygeson A., Jacobsen R.M., Kauserud H., Birkemoe T. (2023) Beetles provide directed dispersal of viable spores of a keystone wood decay fungus. *Fungal Ecology* **63**, 101232: 1–7.
- Majdanová L., Hofmeister J., Pouska V., Mikoláš M., Zíbarová L., Vítková L., Svoboda M., Čada M. (2023) Old-growth forests with long continuity are essential for preserving rare wood-inhabiting fungi. *Forest Ecology and Management* **541**, 121055: 1–9.
- Moor H., Nordén J., Penttilä R., Siitonen J., Snäll T. (2020) Long-term effects of colonization-extinction dynamics of generalist versus specialist wood-decaying fungi. *Journal of Ecology* **109**: 491–503.
- Müller J., Engel H., Blaschke M. (2007) Assemblages of woodinhabiting fungi related to silvicultural management intensity in beech forests in southern Germany. *European Journal of Forest Research* **126**: 513–527.
- Naturalforests.cz (2023) *Proposal for terminology standardization*; https://naturalforests.cz/databank-terminologyproposal-for-terminology (accessed 5 May 2023).
- Nordén J., Harrison P.J., Mair L., Siitonen J., Lundström A., Kindvall O., Snäll T. (2020) Occupancy versus colonization–extinction models for projecting population trends at different spatial scales. *Ecology and Evolution* **10**: 3079– 3089.
- Norros V., Penttilä R., Suominen M., Ovaskainen O. (2012) Dispersal may limit the occurrence of specialist wood decay fungi already at small spatial scales. *Oikos* **121**: 961–974.
- Nuss I. (1999) *Mykologischer Vergleich zwischen Naturschutzgebieten und Forstflächen*. IHW-Verlag, Eching.
- Ódor P., Heilmann-Clausen J., Christensen M., Aude E., Van Dort K.W., Piltaver A., Siller I., Veerkamp M.T., Walleyn R., Standovar T., Van Hees A.F.M., Kosec J., Matocec N., Kraigher H., Grebenc T. (2006) Diversity of dead wood inhabiting fungi and bryophytes in semi-natural beech forests in Europe. *Biological Conservation* **131**: 58–71.
- Ottosson E., Kubartová A., Edman M., Jönsson M., Lindhe A., Stenlid A., Dahlberg A. (2015) Diverse ecological roles within fungal communities in decomposing logs of *Picea abies*. *FEMS Microbiology Ecology* **91**, fiv012: 1–13.
- Ottosson E., Nordén J., Dahlberg A., Edman M., Jönsson M., Larsson K.-H., Olsson J., Penttilä R., Stenlid J., Ovaskainen O. (2014) Species associations during the succession of wood-inhabiting fungal communities. *Fungal Ecology* **11**: 17–28.
- Paillet Y., Bergés L., Hjältén J., Ódor P., Avon C., Bernhardt-Römermann M., Bijlsma R.-J., de Bruyn L., Fuhr M., Grandin U., Kanka R., Lundin L., Luque S., Magura T., Matesanz S., Mészáros I., Sebastià M.-T., Schmidt W., Standdovár T., Tóthmérész B., Uotila A., Valladares F., Vellak K., Virtanen R. (2010) Biodiversity differences between managed and unmanaged forests: meta-analysis of species richness in Europe. *Conservation Biology* **24**: 101–112.
- Parmasto E. (2001) *Fungi as indicators of primeval and oldgrowth forests deserving protection*. In: *Fungal conservation, issues and solutions* (eds. Moore D., Nauta M.M., Evans S.E., Rotheroe M.), Cambridge University Press, Cambridge: 81–88.
- Peay K.G., Bruns T.D. (2014) Spore dispersal of basidiomycete fungi at the landscape scale is driven by stochastic and deterministic processes and generates variability in plantfungal interactions. *New Phytologist* **204**: 180–191.
- Perreaultt L., Forrester J.A., Lindner D.L., Jusino M.A., Fraver S., Banik M.T., Mladenoff D.J. (2023) Linking wood-decay fungal communities to decay rates: Using a long-term experimental manipulation of deadwood and canopy gaps. *Fungal Ecology* **62**, 101220: 1–10.
- Purahong W., Kahl T., Schloter M., Bauhus J., Buscot F., Krüger D. (2014) Comparing fungal richness and community composition in coarse woody debris in Central European beech forests under three types of management. *Mycological Progress* **13**: 959–964.
- Purhonen J., Ovaskainen O., Halme P., Komonen A., Huhtinen S., Kotiranta H., Læssøe T., Abrego N. (2019) Morphological traits predict host-tree specialization in wood-inhabiting fungal communities. *Fungal Ecology* **46**, 100863: 1–8.
- Rajala T., Peltoniemi M., Pennanen T., Mäkipää R. (2012) Fungal community dynamics in relation to substrate quality of decaying Norway spruce (*Picea abies* [L.] Karst.) logs in boreal forests. *FEMS Microbiology Ecology* **81**: 494–505.
- Rajala T., Tuomivirta T., Pennanen T., Mäkipää R. (2015) Habitat models of wood-inhabiting fungi along a decay gradient of Norway spruce logs. *Fungal Ecology* **18**: 48–55.
- Runnel K., Drenkhan R., Adamson K., Lõhmus P., Rosenvald K., Rosenvald R., Rosenvald R., Rähn E., Tedersoo L. (2021) The factors and scales shaping fungal assemblages in fallen spruce trunks: A DNA metabarcoding study. *Forest Ecology and Management* **495**, 119381: 1–10.
- Runnel K., Lõhmus A. (2017) Deadwood-rich managed forests provide insights into the old-forest association of woodinhabiting fungi. *Fungal Ecology* **27**: 155–167.
- Ruokolainen A., Shorohova E., Penttilä R., Kotkova V., Kushnevskaya H. (2018) A continuum of dead wood with vari-

ous habitat elements maintains the diversity of wood-inhabiting fungi in an old-growth boreal forest. *European Journal of Forest Research* **137**: 707–718.

- Rustøen F., Høiland K., Heegaard E., Boddy L., Gange A.C., Kauserud H., Andrew C. (2023) Substrate affinities of wood decay fungi are foremost structured by wood properties not climate. *Fungal Ecology* **63**, 101231: 1–7.
- Seibold S., Müller J., Baldrian P., Cadotte M.W., Štursová M., Biedermann P.H.W., Krah F.-S., Bässler C. (2019) Fungi associated with beetles dispersing from dead wood – let's take the beetle bus! *Fungal Ecology* **39**: 100–108.
- Suominen M., Junninen K., Kouki J. (2019) Diversity of fungi in harvested forests 10 years after logging and burning: polypore assemblages on different woody substrates. *Forest Ecology and Management* **446**: 63–70.
- Šmilauer P., Lepš J. (2014) *Multivariate analysis of ecological data using Canoco 5*. Cambridge University Press, Cambridge.
- ter Braak C.J.F., Šmilauer P. (2012) *Canoco reference manual and user's guide: software for ordination, version 5.0*. Microcomputer Power, Ithaca.
- Tomao A., Antonio Bonet J., Castaño C., de-Miguel S. (2020) How does forest management affect fungal diversity and community composition? Current knowledge and future perspectives for the conservation of forest fungi. *Forest Ecology and Management* **457**, 117678: 1–15.
- Vampola P. (2021) *Poroid fungi (Polyporales s. l.) of Europe in the collections of Muzeum Vysočiny Jihlava*. Nakladatelství Vampola, Jihlava.
- Vampola P., Kunca V., Vlasák J. (2018) Contribution to the knowledge of the rare corticioid species *Steccherinum gracile*. *Mykologické Listy* **140**: 48–56.
- Vrška T., Šamonil P., Unar P., Hort L., Adam D., Král K., Janík D. (2012) *Development dynamics of virgin forest reserves in the Czech Republic*. 3. *Šumava Mts. and Cˇ eský les Mts., Diana, Stožec, Boubín virgin forest, Milešice virgin forest*. Academia, Praha.

(Manuscript accepted 29 January 2024; Corresponding Editor: I. Krisai-Greilhuber)

Naturalness is key: high species richness of wood-inhabiting fungi does not automatically mean high species quality

Electronic Supplements

to paper published in Sydowia (2024)

Jan Holec^{1*}, Pavel Hubený², Tomáš Kučera³

¹ Mycological Department, National Museum, Cirkusová 1740, Praha 9, CZ-193 00, Czech Republic

² Šumava National Park Administration, 1. máje 260, Vimperk, CZ-385 01, Czech Republic

³ Department of Ecosystems Biology, Faculty of Science, University of South Bohemia, Branišovská 1760, České Budějovice, CZ-370 05, Czech Republic

*jan.holec@nm.cz

A. Forest history at Zámecký les (near-natural forest) and localities used for comparison: Boubínský prales (virgin forest) and Mittelsteighütte (natural forest). The colours show the degree of human influence: **green** – almost no human impact, <mark>yellow</mark> – little human impact, red – strong human impact. For sources of data see the main paper: Materials and methods, section Study site Zámecký les and the localities compared.

* state in 2020

B. Position and codes of trunks studied in Zámecký les forest (ZL). **DA**: *Abies alba*, **DP**: *Picea abies*. Source of basic map: Mapy.cz (www.mapy.cz), @Seznam.cz, a.s., 2023.

C. Characteristics of Norway spruce (*Picea abies*) trunks studied in **Zámecký les** (coded DP) and **Boubínský prales** (coded BB). More details on trunks in Boubínský prales, e.g. their identification number in database administrated by the The Silva Tarouca Research Institute for Landscape and Ornamental Gardening (RILOG), Brno, Czech Republic, are available in Holec et al. (2020, 2022).

References

Holec J., Kučera T., Běťák J., Hort L. (2020): Macrofungi on large decaying spruce trunks in a Central European old-growth forest: what factors affect their species richness and composition? – Mycological Progress 19: 53–66.

Holec J., Holcová K., Žák M. (2022): Diversity and ecology of macrofungi on large decaying spruce trunks: what has changed after five years? – Sydowia 75: 23–35.

D. Characteristics of silver fir (*Abies alba*) trunks studied in **Zámecký les** (coded DA) and **Boubínský prales** (coded BA). More details on trunks in Boubínský prales, e.g. their identification number in database administrated by the The Silva Tarouca Research Institute for Landscape and Ornamental Gardening (RILOG), Brno, Czech Republic, are available in Holec & Kučera (2020).

References

Holec J., Kučera T. (2020): Richness and composition of macrofungi on large decaying trees in a Central European old-growth forest: a case study on silver fir (*Abies alba*). – Mycological Progress 19: 1429–1443.

E. Unconstrained ordination biplots (**PCA**) of trunk and habitat variables (**left**) with centroids of trunk positions along the first and second ordinal axes (**middle**), and first and third ordinal axes (**right**). Clumped arrows (highlighted by ellipses) indicate collinear variables. The angles between any two arrows indicate respective pairwise correlations. Negative correlations are denoted by arrows pointing in opposite directions. For codes of environmental variables see Electronic Supplements C, D. Nspec represents species richness. The folded aspect of fallen trunk to south-west (FAsw) is based on azimuth value and computed as described in Holec et al. (2019).

Reference

Holec J., Běťák J., Dvořák D., Kříž M., Kuchaříková M., Krzyściak-Kosińska R., Kučera T. (2019) Macrofungi on fallen oak trunks in the Białowieża Virgin Forest – ecological role of trunk parameters and surrounding vegetation. Czech Mycology 71: 65–89.

F. Fungal species recorded on the studied **Norway spruce (***Picea abies***)** trunks in Zámecký les (this study, years 2021–2022) and Boubínský prales (data from Holec et al. 2022, year 2020). Cumulated presence/absence data from 4 visits per trunk. Categories of the Czech Red List (Holec et Beran 2006) plus rarity and specificity of some species are indicated in third column (SSI species). For details on trunks (Zámecký les: coded DA, Boubínský prales: coded BA) see Electronic Supplements B, C. Species written in **black** occur at both localities, in **red** only in Zámecký les, in **blue** only in Boubínský prales in 2020, in **violet** in Zámecký les and also in Boubínský prales in 2015 (data from Holec et al. 2020).

Abbreviations:

* generally rare and old-growth forests species

^A as *B. vagum* in Holec & Kučera (2020)

^B *Botryobasidium* sp. from 4-species group sensu Bernicchia & Gorjón(2010)

^C as *Tulasnella inclusa* in Holec et al. (2022)

cr: critically endangered

dd: data deficient

en: endangered

nt: near threat

vu: vulnerable

References

Bernicchia A., Gorjón S.P. (2010) Corticiaceae s.l. – Edizioni Candusso, Alassio.

Holec J., Beran M., eds. (2006): Červený seznam hub (makromycetů) České republiky [Red list of fungi (macromycetes) of the Czech Republic]. – Příroda, Praha, 24: 1– 282.

Holec J., Holcová K., Žák M. (2022) Diversity and ecology of macrofungi on large decaying spruce trunks: what has changed after five years? – Sydowia 75: 23–35. Holec J., Kučera T. (2020) Richness and composition of macrofungi on large decaying trees in a Central European old-growth forest: a case study on silver fir (Abies

alba). – Mycological Progress 19: 1429–1443.

Holec J., Kučera T., Běťák J., Hort L. (2020) Macrofungi on large decaying spruce trunks in a Central European old-growth forest: what factors affect their species richness and composition? – Mycological Progress 19: 53–66.

G. Fungal species recorded on the studied Silver fir (Abies alba) trunks in Zámecký les (this study, years 2021–2022) and Boubínský prales (data from Holec & Kučera 2020, years 2017-2019). Cumulated presence/absence data from 4 visits per trunk. Categories of the Czech Red List (Holec et Beran 2006) plus rarity and specificity of some species are indicated in third column (SSI species). For details on trunks (Zámecký les: coded DA, Boubínský prales: coded BA) see Electronic Supplements B, C. Species written in black occur at both localities, in red only in Zámecký les, in blue only in Boubínský prales.

Abbreviations:

* generally rare and old-growth forest species

§ species preferring wood of Abies

^A as *B. vagum* in Holec & Kučera (2020)

^B Botryobasidium sp. from 4-species group sensu Bernicchia & Gorjón(2010)

 c ex aff. Botryobasidium simile in Holec & Kučera (2020) = B. ex aff. conspersum sensu Zíbarová (2021)

^D as Tomentella stuposa in Holec & Kučera (2020)

E as Tomentella sp. sect. Alytosporium in Holec & Kučera (2020)

cr: critically endangered

dd: data deficient

en: endangered

nt: near threat

vu: vulnerable

References

Bernicchia A., Gorjón S.P. (2010) Corticiaceae s.l. - Edizioni Candusso, Alassio.

Holec J, Beran M., eds. (2006) Červený seznam hub (makromycetů) České republiky [Red list of fungi (macromycetes) of the Czech Republic]. – Příroda, Praha, 24: 1– 282.

Holec J., Kučera T. (2020) Richness and composition of macrofungi on large decaying trees in a Central European old-growth forest: a case study on silver fir (Abies alba). - Mycological Progress 19: 1429-1443.

Svrček M. (1960) Tomentelloideae Čechoslovakiae, Genera resupinata familiae Thelephoraceae s.str. - Sydowia 14: 170-245.

Zíbarová L. (2021) Pavučiník podobný – Botryobasidium simile nalezen i v Čechách (Botryobasidium simile – first record in Bohemia). – Mykologické Listy 150: 39–48.

H. The most frequent macrofungi on studied trunks of *Picea abies* in Zámecký les (ZL).

The species occur on 6–9 trunks of 12 studied.

I. The most frequent macrofungi on studied trunks of *Abies alba* in Zámecký les (ZL).

The species occur on 4–5 trunks of 6 studied.

